

*Full Length Research Paper*

# Heavy metals concentrations and burden in the bivalves (*Anadara (Senilia) senilis*, *Crassostrea tulipa* and *Perna perna*) from lagoons in Ghana: Model to describe mechanism of accumulation/excretion

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Seasonal sampling of the bivalves: *Anadara (Senilia) senilis* (n = 260), *Crassostrea tulipa* (n = 220), from two 'open' lagoons (Benya and Ningo) and a 'closed' lagoon (Sakumo), and *Perna perna* (n = 170), from rocky shores adjacent to Benya and Sakumo, were analyzed for their total Cu, Zn, Fe, Mn, Cd and Hg concentrations and total body burden (that is concentration multiply by total flesh weight) . Median concentrations for cockles were: 5, 38, 710, 10, 0.22 and 0.20 µg/g dw respectively. Cd and Hg levels in dry season samples were higher than those in wet season. While Zn and Fe dry season levels were lower than wet season with no variation in Cu and Mn. Median total body burden in cockles were: 3.3, 30.5, 370, 5.2, 0.28 and 0.13 µg respectively. Geographical variations observed were mostly due to size distribution rather than the ecological differences between stations. Log-transformed regression relationships between body burden and size were used to compare the species. Relationship between condition index, body burden and concentration were used to develop a model to describe mechanism of accumulation/excretion.

**Key words:** Heavy metals, bivalves, marine, lagoons, seasonal variation, accumulation, Ghana.

## INTRODUCTION

Monitoring programs and research for metals in the environmental samples have become widely established because of concerns over accumulation and toxic effects, particularly in aquatic organisms and to humans consuming these organisms. The criteria by which organisms are accepted as biological indicator for the assessment of contamination were proposed more than twenty five years ago and remain unchanged (Phillips, 1976; Fowler and Oregioni, 1976). Bivalves are widely used as bioindicators of heavy metals pollution in coastal areas because they are known to concentrate these elements, providing a time integrated indication of environmental contamination. In comparison to fish and

crustacea, bivalves have a very low level of activity of enzyme systems capable of metabolizing persistent organic pollutants (POPs), such as aromatic hydrocarbons and polychlorinated biphenyls. Therefore contaminants concentrations in the tissues of bivalves more accurately reflect the magnitude of environmental contamination (Phillips, 1977, 1980, 1990).

Factors known to influence metal concentrations and accumulation in these organisms include metal bioavailability, season of sampling, hydrodynamics of the environment, size, sex, changes in tissue composition and reproductive cycle (Boyden and Phillips, 1981). Seasonal variations have been related to a great extent

to seasonal changes in flesh weight during the development of gonadic tissues (Joiris et al., 1998, 2000). Element concentrations in molluscs at the same location differ between different species and individuals due to species-specific ability/capacity to regulate or accumulate trace metals (Reinfelder et al., 1997; Otchere et al., 2003). Different animals in the same community at the same trophic level could accumulate pollutants differently due to differences in habitat/niche's physical and chemical properties.

This paper presents data on the total concentration and body burden of six metals, the effect of season on these bivalves and a model developed to explain excretion and accumulation of these metals in the bivalves.

## MATERIALS AND METHODS

Three species of bivalves were collected from the lagoons of Ghana (which lies between latitude 4° and 115° N) with approximately 590 km long coastline, stretches from 3°W to 1° 10'E and lies between 4°5' and 6°6'N. About 50 lagoons occur on the coast. The lagoons are of two main types: 'open' and 'closed' lagoons. The open lagoons are in contact with the sea throughout the year and therefore partly under tidal influence. Temperature and salinity ranges are 24 to 32°C and 10 to 40 psu, respectively. The closed lagoons are cut off from the adjacent sea by a sand bar (about 40 m wide) for the greater part of the year. Temperature and salinity ranges are: 27 to 34°C and 27 to 70 psu, respectively. Hyperhaline condition results from evaporation during the dry season (Yankson, 1982).

Samples of *Anadara (Senilia) senilis* (cockles; n = 260), and *Crassostrea tulipa* (oysters; n = 220) from three lagoons and *Perna perna* (mussels; n = 170) from two rocky shores adjacent to these

lagoons, were collected in October 1996 and February 1997 (wet and dry seasons, respectively). The analytical procedure used to measure the metals Cd, Cu, Fe, Mn and Zn; in the bivalves was based on UNEP/FAO/IAEA (1982) with modification as follows: sub-sample (dried) tissue (0.2 g) was heated with 10 ml of concentrated nitric acid (70 to 90°C) till all tissue had been digested. The temperature was then gradually increased to 135 °C

and drops of H<sub>2</sub>O<sub>2</sub> added for further oxidation. After cooling, solutions were diluted to 50 ml with double distilled water and filtered with 1.6µm fiberglass filter paper (GF/A).

Samples were then stored at room temperature in 50 ml volumetric flask until they were analyzed. Analyses were carried out with flame Atomic Absorption Spectrophotometer (AAS) (Perkin Elmer 3110 system 2) with electrode discharge lamps (EDL) and hollow cathode lamp (HCL). Wavelengths and detection limits are shown in Table 2. Air-acetylene oxidizing flame was used, with 1ml autosampler, each sample was measured three times and the concentration of the sample given as the mean of the three measurements. The AAS was calibrated for each metal with the following standard solutions 5, 10, 15, 20 and 25 ppm (from 1000 ppm stock solution) before each metal determination. All reagents used were of analytical grade (Analar). Method for the determination of Hg has been described in Joiris et al. (1998). Samples prepared for metal analysis included procedural blanks, replicate analyses, standard solutions and certified reference material (CRM) of dogfish muscle (DORM-2, National Research Council of Canada, Marine Analytical Chemistry Standards Program) . The certified values for selected heavy metals are shown in Table 1 (no corrections carried out).

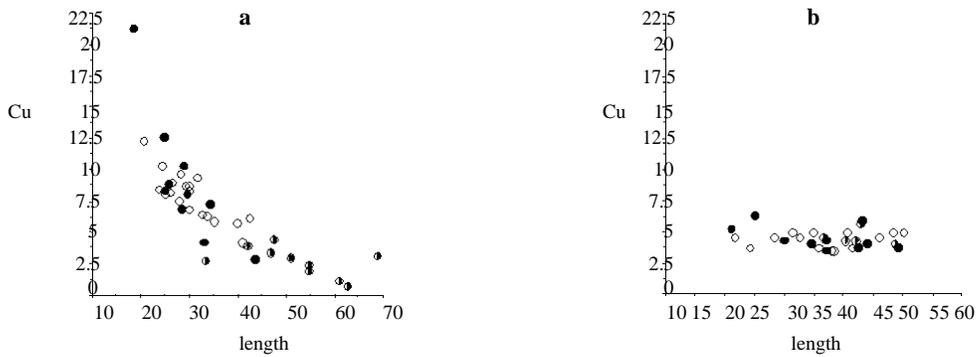
The data were expressed as total concentration and body burden. Median values are presented and differences were tested using the Mann-Whitney U-test of significance due to the non-normal distribution of the data. The species were compared using log transformed regression analysis. Relationship between condition index, body burden and concentration were used to develop a model to describe mechanism of accumulation/excretion.

**Table 1.** Comparison between certified metal concentrations in CRM DORM-2 and values obtained from this study (µg/g dw).

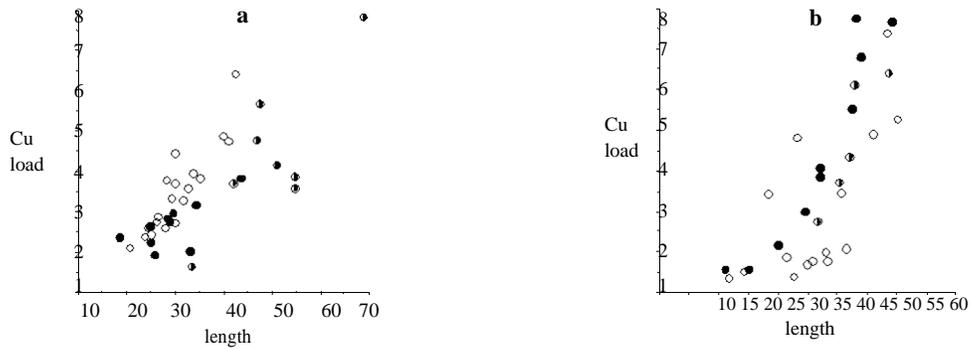
Element	Certified value	Value found	% of target value
Cadmium	0.043 ± 0.008	0.036 ± 0.006	84
Copper	2.34 ± 0.16	2.04 ± 0.16	87
Iron	142 ± 10	133 ± 14	94
Manganese	3.66 ± 0.34	3.2 ± 0.12	87
Mercury	0.789 ± 0.074	0.765 ± 0.029	96
Zinc	25.6 ± 2.3	22.7 ± 2	89

**Table 2.** Wavelengths (nm) used to measure the metals, sensitivity of the AAS (mg/l) and detection limits (ng/g).

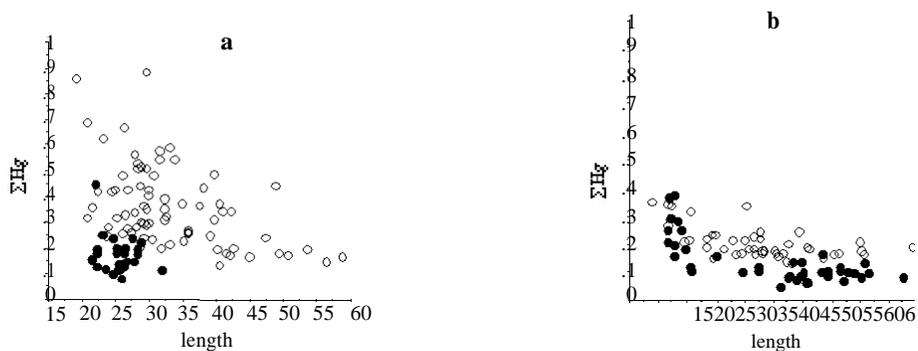
Element	Wavelength	Sensitivity	Detection limit
Cadmium	228.8	0.0005	2.13
Copper	324.7	0.001	3.9
Iron	248.3	0.003	183
Manganese	279.5	0.001	30
Mercury	253.7	0.001	40
Zinc	213.9	0.0008	441



**Figure 1.** Relationship between Cu concentrations ( $\mu\text{g/g dw}$ ) in cockles and size (mm): dry (a) and wet (b) seasons; Benya (circles), Ningo (half shaded) and Sakumo (dots) lagoons.



**Figure 2.** Relationship between Cu load ( $\mu\text{g}$ ) and length (mm) of *A. senilis*, dry (a) and wet (b) seasons; Benya (circles), Ningo (half shaded) and Sakumo (dots) lagoons.



**Figure 3.** Relationship between  $\Sigma\text{Hg}$  concentrations ( $\mu\text{g/g dw}$ ) in cockles and size (mm): dry (a) and wet (b) seasons; Benya (circles), Ningo (half shaded) and Sakumo (dots) lagoons.

## RESULTS

Median concentrations and body burden for both seasons are shown in Table 3. One important factor influencing these results is size. For example, Cu concentration in

the dry season (cockles) decreases with size as a "growth dilution" effect but stable in the wet season (Figures 1a and b). In this case regional differences in Cu could be entirely due to size effect. Likewise, the seasonal difference in Cd and Hg dry season

**Table 3.** Seasonal and locational variation in concentration ( $\mu\text{g/g dw}$ ), load (burden,  $\mu\text{g}$ ) of total Cu, Zn, Fe Cd, Hg, Mn and size (mm) in the bivalves studied: median values; Benya and Ningo are open lagoons and Sakumo a closed lagoon n = number of samples.

	n	Size	Concentration						Body burden					
			Cu	Zn	Fe	Mn	Cd	Hg	Cu	Zn	Fe	Mn	Cd	Hg
<b>Dry season</b>														
<i>Anadara senilis</i>														
Benya	20	29	8.2	36	790	5	0.90	0.36	3.4	13	350	2.7	0.35	0.17
Ningo	10	53	3.0	36	210	15	0.34	0.20	3.8	51	370	26	0.53	0.34
Sakumo	10	29	8.2	6	520	18	1.1	0.16	2.7	1.5	150	4.3	0.35	0.05
<i>Crassostrea tulipa</i>														
Benya	20	47	17	2350	350	11	0.74	0.21	9	1290	230	7.2	0.41	0.15
Ningo	10	42	74	380	450	20	1.1	0.16	31	160	180	8.0	0.44	0.13
Sakumo	10	45	40	430	280	17	0.91	0.13	19	290	130	7.5	0.42	0.15
<i>Perna perna</i>														
Benya	20	40	15	16	900	12	1.4	0.37	3.6	4.3	230	3.3	0.36	0.08
Sakumo	10	39	16	12	1130	15	1.9	0.33	3.0	2.0	220	2.9	0.36	0.19
<b>Wet season</b>														
<i>Anadara senilis</i>														
Benya	20	36	4.6	104	1100	10	0.30	0.21	2.0	57	530	4.0	0.15	0.11
Ningo	5	42	4.3	43	830	8	0.19	0.14	4.4	43	830	7.3	0.16	0.14
Sakumo	10	37	4.5	35	570	19	0.13	0.11	4.0	26	350	15.9	0.16	0.13
<i>Crassostrea tulipa</i>														
Benya	15	32	17	2780	560	13	0.21	0.14	5	600	130	2.8	0.08	0.04
Ningo	10	32	59	560	700	13	0.18	0.13	12	120	140	3.1	0.05	0.07
Sakumo	10	43	33	660	520	20	0.12	0.12	19	250	260	11.0	0.09	0.10
<i>Perna perna</i>														
Benya	15	40	8.4	87	1050	18	0.42	0.20	2.1	27	290	5.1	0.12	0.05
Sakumo	10	38	7.0	47	610	5	0.11	0.26	1.4	11	120	1.0	0.02	0.04

**Table 4.** Slope coefficient (b), intercept (a) and proportion of variation explained ( $r^2$ ) of the metal burden and the size of bivalves studied (both seasons and all stations pooled).

Metal	Cockles			Oysters			Mussels		
	a	b	r	a	b	r	a	b	r
Copper	-1.01	0.99	0.39	-0.95	1.26	0.19	-4.2	2.89	0.18
Zinc	-3.88	3.37	0.60	0.77	1.16	0.10	-1.99	1.74	0.02
Iron	0.30	1.48	0.37	0.80	0.9	0.20	-1.05	2.12	0.17
Cadmium	-4.83	2.57	0.44	-3.61	1.46	0.29	-6.31	3.28	0.06
Manganese	-3.4	2.73	0.50	-1.34	1.31	0.31	-5.82	3.93	0.28
Mercury	-3.36	1.62	0.48	-3.18	1.34	0.05	-3.74	1.64	0.09

concentrations were also apparently higher than those in the wet season. A closer look at the total body burden data with respect to these metals seasonal differences could be detected. Comparing Benya and Sakumo wet season body burden (with larger cockle sizes; 37 mm) to dry season (29 mm) suggest a lower metal availability (Cd) at the stations during the wet season. Thus the

above seasonal differences in Cd and Hg concentrations are real.

Both body burden and concentrations in these metals showed no geographical difference during the seasons and any seeming difference was due to size/age. While Zn, Fe and Mn dry season concentrations were lower than wet season. Again considering the body burden the

**Table 5.** Slope coefficient (b), intercept (a) and proportion of variation explained ( $r^2$ ) of the metal load and the size of cockles; dry season (see legend to table 3).

Metal	Benya			Ningo			Sakumo		
	a	b	$r^2$	a	b	$r^2$	a	b	$r^2$
Copper	-1.49	1.37	0.82	-1.12	1.0	0.25	-0.4	.57	0.39
Zinc	-3.3	3.01	0.55	-1.75	2.05	0.88	-4.94	3.63	0.79
Iron	-.34	1.95	0.90	-0.53	1.79	0.56	-0.72	2.04	0.77
Cadmium	-5.54	3.34	0.76	-3.99	2.17	0.85	-1.05	0.39	0.25
Manganese	-2.94	2.24	0.43	-1.21	1.48	0.14	0.41	0.15	0.01
Mercury	3.48	1.83	0.46	1.47	0.62	0.29	3.66	1.7	0.19

$Y = aW^b$ , where Y is the load, W the size, a the intercept and b the slope co-efficient (Boyden, 1974).

influence of size is obvious from the table; Mn did not show any difference both in location and season. Any detected differences in concentration and burden were apparent. On the other hand seasonal differences could be found in Fe and Zn concentrations. For example, Ningo dry season burden (with higher biomass) was lower than wet season, an indication of a higher contamination of these metals in the wet season. Geographical differences could be detected in other species, such as higher Zn concentration in oysters from Benya during both seasons. Seasonal variations on the other hand, could be detected mainly due to hydrological differences like the effect of rain or differences between open and closed lagoons. Elevated levels of Zn in oysters (270 - 3540  $\mu\text{g/g dw}$ ), Fe in cockles (150- 2050  $\mu\text{g/g dw}$ ) and in mussels (510 - 1880  $\mu\text{g/g dw}$ ) reflect the inherent variability of metals in bivalves dependence on the particular species-metal pair considered, and also on the degree of contamination involved (Boyden and Phillips, 1981).

## DISCUSSION

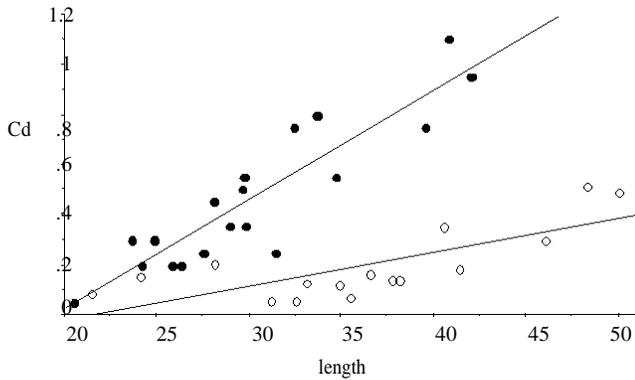
In order to avoid the problem of cases where it is difficult to assess whether observed differences in tissue level/burden between these bivalves reflect real differences in environmental heavy metal constitution, or are merely due to variations in body size, plots of  $\mu\text{g}$  metal per individual (total burden versus size) was carried out. Generally, the total metal burden is related to body

weight/size by the metabolic power function:  $Y = aW^b$ , where Y is the burden, W the dry weight, a the intercept and b the slope co-efficient (Boyden, 1974). Coefficients  $> 2$  of the body weight as shown by Cd, Zn and Mn in cockles and Cu, Fe, Cd and Mn in mussels (Table 3) can best be explained as being due to removal of this elements from body circulation and accumulation within specific tissues, possibly as a result of some exceptional affinity. In the case of the oysters (Table 4) all the relationships were approximately equal to one ( $b = 1$ ), implying that element concentration was independent of size. The absolute amount of the metal is not dictated by

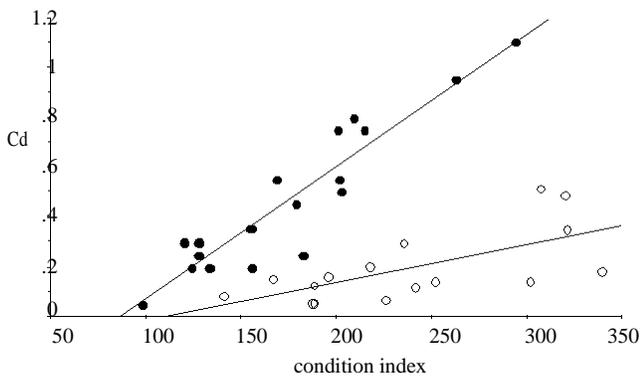
the amount of such binding compound/storage mechanisms in the tissues.

The tendency to decrease the metal concentrations with an increase in body weight is not significant in all cases. However, for the essential metals Cu and Fe (in cockles and oysters) the size of the animal could be mainly responsible for the metal concentration variability. The slope value of about 2 in cockle population examined (Table 5) may well be due to the unusually elevated environmental Fe concentration in the lagoons. An increase in metal burden with size/weight of the individuals is expected and it is indeed noted for all metals, except for Zn and Cd in mussels. Information about the factors responsible for the observed metal burden-size can be obtained from the slope of the regression lines (Table 5).

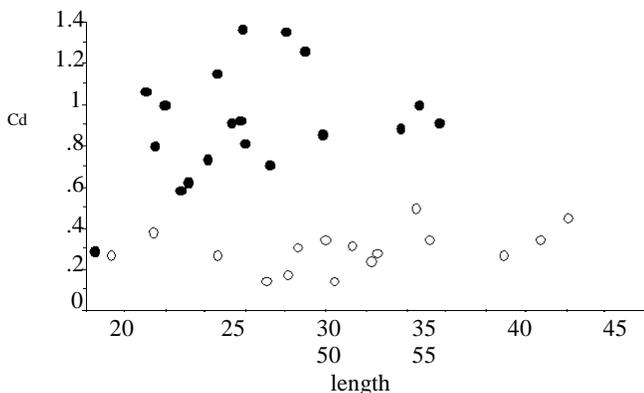
The changes in metabolic rates of bivalves with length and season as well as the variation in bioavailability of metals in the surrounding environment with time might be responsible for these variations in these molluscs (Boyden, 1974; Cossa and Rondeau, 1985). Metal availability probably fluctuated with salinity in these lagoons during the seasonal cycle (Phillips, 1977; Cossa and Rondeau, 1985). This probably caused the higher metal burden and concentration in the wet season (such as for Fe and Zn). Higher wet season levels of Fe and Zn might as well be due to 'import' as most roofing in Ghana are made of galvanized iron sheets. Lower wet season levels for Cd and Hg could be attributed to lost through spawning or washed-out of the lagoons during the rainy period. Many metals are also found in agricultural products. Those present in fertilizers include Cd, Cu, Cr, Ni, Mn, Mo and Zn. Eventually, many of these metals may accumulate in soils and become exposed to run-offs during the rainy season. Biological variables such as size, sex or changes in tissue composition and reproductive cycle as well as the season of sampling and the hydrodynamics of the lagoons have to be considered. Seasonal variations have been reported to be higher in winter/dry than in summer/wet. These seasonal variations have been related to a great extent to seasonal changes in flesh weight during development of gonadic tissues (Cossa and Rondeau, 1985; Joiris et al., 1998; Otchere et al., 2000, 2003).



**Figure 4.** Relationship between Cd body burden ( $\mu\text{g}$ ) and size (mm) of cockles from Benya; dry season (dots) and wet season (circles).



**Figure 5.** Relationship between total Cd body burden ( $\mu\text{g}$ ) and CI of cockles from Benya; dry season (dots) and wet season (circles).



**Figure 6.** Relationship between Cd concentration ( $\mu\text{g/g dw}$ ) and size (mm) of cockles from Benya; dry season (dots) and wet season (circles).

Metabolic changes of the mollusc with age and season, storage mechanisms as well as temporal variations in metal bioavailability in the surrounding environment might be the reasons to explain these burden-size relationships in the lagoons (Tables 4 and 5). In molluscs, absolute

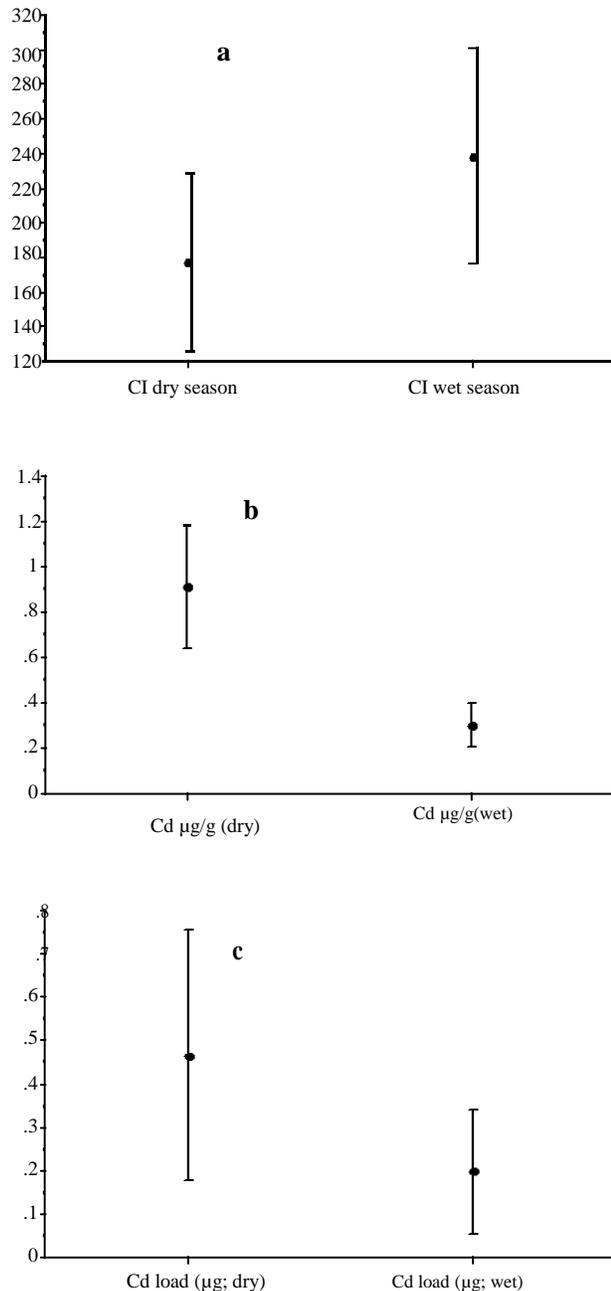
tissue levels are related or determined by environmental levels. In cases where the body burden is directly related to body weight, as has been found for a variety of elements in several species (Boyden, 1974), a function of body weight such as binding of specific compounds within the tissue may play some role in determining the total burden. The amount of metal, however, may not depend on the amount of such binding compounds within the tissues but on the amount of metal available in that environment. Thus both biotic and abiotic factors probably constantly change their relative importance in these bivalves and it is unlikely that a steady state could ever be reached. Finally seasonal and regional variations observed in this study were likely to be due to these factors mentioned above. Results on metal contamination in bivalves should therefore not to be used directly as reflecting environmental contamination; they should first be normalized for these physical and environmental factors before any spatial and/or temporal comparison is made.

#### Excretion and accumulation model

Metal penetration in bivalves may take place in several ways: diffusion of ions or complexes, mediated transport and/or endocytosis of particulate metal and pinocytosis of organo-metallic aggregates. Uptake may take place at the gill or digestive gland or on the surface of the mantle, and their relative importance is a function of the speciation of the metal in the environment. Differences in metal load (body burden) and flesh weight of the bivalves have been the main influences on the level of contamination in this study. The origin and range of variations in concentrations of these pollutants can be described through the Figures 4 to 7.

Considering Cd concentration in cockles in the wet and dry seasons, the slope of the Cd/size relationship results from a combination of several factors which act separately or simultaneously. The figures reflect the competition between the opposite effect of age and growth on the accumulation of Cd. The metal accumulates with age whereas the weight gain of the individual has the effect of reducing the metal concentration. The relationship observed underlying the linear relation could effectively correspond in a given range of soft tissue weights to the presence of individuals of the same age class whose soft tissue weight differs considerably. In this way, while the Cd burden would increase with age (Figure 4), concentration would diminish with the weight/size (Figure 6) within each age class if growth is more rapid than the accumulation rate.

To the inverse effect of age and soft tissue weight is added the effect of change in the bioavailability of Cd in the environment. To demonstrate these mechanisms clearly the phenomenon in the light of situations encountered in cockles will be restated: that in bioaccumulation in bivalves, the uptake of the metal is higher in the smaller individual than in the larger ones generally resulting in "growth dilution" effect. The result is



**Figure 7.** One bar standard error from the mean for CI (a), body burden (b) and concentrations (c) in cockles from Benya during the dry and wet seasons.

a negative slope for metal/size concentration relationships in the case of Hg (Figure 3) but not in Cd as in Figure 6 (for the non-essential metals), which was only due to seasonal effect. Higher metabolism in juvenile individuals can cause a reduction in metal concentrations in soft tissue, because the tissue grows more quickly than the metal can be absorbed. This mechanism can explain seasonal variations recorded in Cd concentrations. In Figure 7, condition index (CI) in the wet season was

higher than the index in the dry season while load was higher in the dry season than in the wet season expressing their opposite effect on concentration and explaining the seasonal variations observed in Cd concentration in cockles.

Applying the same principle as in Cd to Zn and Cu would explain the variations observed. CI was higher in the wet season than in the dry season for both metals. For example, while load (body burden) was higher in dry season for Cu expressing the opposite effect on concentration to explain the temporal difference observed. Comparative differences between wet and dry seasons for Cu and Zn were about 1.3-fold and 3-fold increase respectively. The difference was not significant in Cu because as an essential element it is strictly regulated; contrarily Zn showed a significant seasonal variation due to the elevated level in the Ghana environment as a result of galvanized roofing sheet used thus Zn becomes more available during the rainy period than the dry season. Therefore Zn regulation in cockles could be described to have broken down due to elevated levels in the environment. It is clear that concentrations of all 3 metals were mainly determined by cockle condition and seasonal fluctuations in metal concentration, which in turn depended to a large extent on the magnitude and direction of temporal alteration in the dry weight of whole soft tissue of the cockles.

Fowler and Oregioni (1976), Phillips (1976) and Boyden and Phillips (1981) studying *Mytilus galloprovincialis*, *M. edulis* and *Crassostrea gigas* respectively, concluded that temporal variations in metal concentrations were mainly caused by changes in soft tissue weights of the bivalves according to the sexual cycle. Thus alterations in the concentrations of metals reciprocated those of the whole soft tissue weights. Investigations of element seasonality in bivalves molluscs are important in terms of their implications for the use of these organisms as biological indicators of metal abundance. The present study suggest that element concentrations may commonly vary in *Anadara senilis* by a factor of 2 or more during the annual cycle. Clearly, such variation is a potential disruptive factor in the use of *A. senilis* to monitor metal abundance at different locations and seasons. The simultaneous sampling of cockles at all sites in such an indicator survey might be considered sufficient to overcome this source of variations, but even in this case there can be no guarantee that cockles from each locations are of similar condition. The best method for diminishing the effects of seasonality is therefore to time sampling programs to occur during a period of relatively little change in element concentrations. In addition, the effect of season could be eliminated by the removal of gonads prior to analysis.

## Conclusion

Concluding, the wet season maxima in Zn, Fe and Mn observed should reflect a higher metal availability during this season (through 'import'). The four essential metals

which were studied, were present in similar respective concentrations to those found for other bivalves elsewhere and exhibited similar seasonal pattern in terms of their concentrations although of different magnitudes (Boyden and Phillips, 1981; Reinfelder et al., 1997; Joiris and Azokwu, 1999). There was no influence of season and location on Mn and Cu concentrations, but location played an important role in Fe and Zn concentrations while season exhibited a moderate influence. Temporal variations in metal concentrations were mainly caused by changes in soft tissue weights of the bivalves according to the sexual cycle. Thus alterations in the concentrations of metals reciprocated those of the whole soft tissue weights hence its respective effect on excretion and accumulations.

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